



Contents lists available at ScienceDirect

Marine Pollution Bulletin

journal homepage: [www.elsevier.com/locate/marpolbul](http://www.elsevier.com/locate/marpolbul)

## A comparative study of marine litter on the seafloor of coastal areas in the Eastern Mediterranean and Black Seas

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### ARTICLE INFO

#### Article history:

Available online xxxxx

#### Keywords:

Litter  
Benthic  
Pollution  
Eastern Mediterranean  
Black Sea

### ABSTRACT

In the present work, abundance, spatial distribution and qualitative composition, of benthic marine litter, were investigated in five study areas from the Eastern Mediterranean and Black Seas (Saronikos, Patras and Echinades Gulfs; Limassol Gulf; Constanta Bay). Surveys were performed using the monitoring protocol proposed by the Technical Group for Marine Litter. Densities ranged from 24 items/km<sup>2</sup> to 1211 items/km<sup>2</sup>, with the Saronikos Gulf being the most affected area. *Plastics* were predominant in all study areas ranging from 45.2% to 95%. *Metals and Glass/Ceramics* reached maximum values of 21.9% and of 22.4%. The size distribution of litter items showed that  $\geq 50\%$  fall into medium size categories (10 × 10 cm, 20 × 20 cm) along with an elevated percentage of small-sized (<5 × 5 cm) plastic litter items. The comparative analysis of the data highlighted the dependence of the marine litter problem on many local factors (human sources and oceanographic conditions) and the urgent need for specific actions.

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### 1. Introduction

In 2012, global production of waste had reached 3.4 million tons and this figure is expected to double by 2025, while about half of this amount concerns non-biodegradable materials (i.e. plastics and metals) (World Bank, 2012). It is an evident fact that much of this waste escapes management schemes and ends up in the environment. More specifically, in the marine environment litter can be transported from land, via rivers, stormwater, wind and sewage, or can be disposed off directly at beaches and at sea (UNEP, 2011). The growing amounts of generated litter and the slow degradation rates result in the accumulation of litter in the oceans and pose a serious threat for healthy oceans and indeed, litter has been found widespread in the global marine environment (Barnes et al., 2009).

The impacts of marine litter on marine life and the environment include deviation in epi-benthic community structure and composition (Katsanevakis et al., 2007); entanglement of organisms

especially in derelict fishing gear and 'ghost nets' (Waluda and Staniland, 2013; Adimey et al., 2014; Anderson and Alford, 2014) and ingestion (Goldstein and Goodwin, 2013; Lusher et al., 2013; Anastasopoulou et al., 2013; Bond et al., 2013; Madeira Di Benedetto and Arruda Ramos, 2014). In addition floating litter facilitates the migration of non-native species to other marine regimes. Also it has been demonstrated that microplastic fragments are sources of persistent organic pollutants (POPs) such as phthalates (Lopez-Carillo et al., 2010; Mathalon and Hill, 2014) and have the potential to concentrate POP's on their surface and consequently act as pollutant carriers (Karapanagioti and Klontza, 2007, 2008; Karapanagioti et al., 2011; Bakir et al., 2014; Mizukawa et al., 2013; Ivar do Sul and Costa, 2014) with secondary effects to organisms which eventually digest them.

Benthic litter tend to become trapped in areas of low circulation and high sediment accumulation in contrast to floating litter, which accumulates in frontal areas. Litter that reaches the seabed may already have been transported considerable distance, only sinking when weighed down by entanglement and fouling. The consequence is an accumulation of litter on specific seabed locations in response to local sources and oceanographic conditions

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(Galgani et al., 2000; Keller et al., 2010; Watters et al., 2010; Ramirez-Llodra et al., 2013; Pham et al., 2013).

Marine litter has been an issue of concern by the United Nations Environment Programme since the early 1970s (UNEP, 2011). In Europe, Marine Strategy Framework Directive (MSFD) (2008/56/EC) has been developed in order to protect the marine environment as well as to ensure its sustainable use. The ultimate goal of MSFD is the achievement of Good Environmental Status (GES) for the marine environment and the *Clean Seas by 2020*, by the Member States. MSFD is based on eleven Descriptors and sets targets for each Descriptor that should be met in order to achieve the GES. Marine litter is addressed as Descriptor 10: “*Properties and quantities of marine litter do not cause harm to the coastal and marine environment*”.

A first step to understanding and alleviating the problem of marine litter is to quantify the existing amounts. Especially for the implementation of the MSFD, this information is urgently needed for setting the targets of Descriptor 10 – Marine Litter. Information on marine litter quantities and distribution in the Mediterranean Sea is still fragmented and its impacts on the marine ecosystem have not been properly addressed. To date, systematic studies dealing with benthic litter have focused on the western Mediterranean Sea (Galgani et al., 1995, 1996, 2000; Ramirez-Llodra et al., 2013; Pham et al., 2013) while for the Eastern Mediterranean Sea relevant information covers only few geographical areas (Galil et al., 1995; Koutsodendritis et al., 2008; Stefatos et al., 1999; Katsanevakis and Katsarou, 2004; Sanchez et al., 2013). In the Black Sea, only one reported work deals with the occurrence and distribution of benthic litter on the western continental shelf in the vicinity of Bosphorus Straits (Topçu and Öztürk, 2010). Therefore, more efforts are needed in order to resolve many scientific questions and to provide the stakeholders on the regional and national levels, with the appropriate information and knowledge.

The objective of this study is to determine the abundance, spatial distribution and qualitative composition of benthic marine litter as well as to investigate the factors affecting benthic litter accumulation and distribution (external sources vs internal transport mechanisms) in five selected study areas in the Eastern Mediterranean (Saronikos, Echinades, Patras Gulfs in Greece, Limassol Gulf in Cyprus) and Black Seas (Constanta Bay in Romania). The monitoring protocol proposed by the Technical Group on Marine Litter was applied throughout our surveys (Galgani et al., 2013b). Up to our knowledge, this is the first comparative study on benthic marine litter including several coastal areas in the interconnected Eastern Mediterranean and Black Seas. In particular for three of our study areas (Saronikos Gulf, Limassol Gulf, Constanta Bay), the present work provides the first recorded data on benthic marine litter.

## 2. Study areas

Benthic marine litter was investigated in the Saronikos Gulf (SR), the Gulf of Patras (PT), the Echinades Gulf (EC) in Greece; the Limassol Gulf (LM) in Cyprus and in the Constanta Bay (CN) in Romania (Fig. 1).

Saronikos Gulf is located in the South Aegean Sea. It is a semi-enclosed gulf (2600 km<sup>2</sup>) which practically constitutes the sea border of the metropolitan city of Athens and the alongshore outskirts (Attica region; approx. 4 million inhabitants). The Piraeus city-port (~17,525 arrivals in 2013; [www.olp.gr](http://www.olp.gr)) is located at its northeastern edge. Saronikos Gulf is characterized by extreme marine navigation, tourism and well-developed fisheries (professional and recreational). Our samplings covered the western basin of the gulf, which is basically an elongated north–south trough

with maximum depths of ~220 m in the north and ~450 m in the south consisting the deepest part of the gulf.

The Gulf of Patras and the Echinades Gulf are interconnected and located in the Ionian Sea. The Echinades Gulf is formed by the Greek mainland and the islands of Kefalonia and Ithaki with maximum water depth of about 350 m. The Echinades Gulf connects the Gulf of Patras with the Ionian Sea. The coastal area of the gulf is used for agriculture without important population nuclei. Approximately 150 shipping routes are recorded weekly, connecting the city of Patras to the Ionian Islands and Italy. Moreover, the gulf is considered as one of the most important fishing grounds and aquaculture of the Ionian Sea. A major river, Acheloos River, flows to the Echinades Gulf with an average annual flow of about  $7800 \times 10^6$  m<sup>3</sup> of water. The Gulf of Patras constitutes a relatively shallow (max 150 m depth), semi-enclosed gulf (400 km<sup>2</sup>), to the east of the Echinades Gulf. The coasts around the Gulf of Patras are urbanized (the city of Patras hosts ~250,000 inhabitants); marine navigation (Patras port ~1923 arrivals in 2013; [www.patrasport.gr](http://www.patrasport.gr)), tourism and well-developed professional and recreational fisheries are the major anthropogenic activities in the gulf. River runoffs (Evinos River, Pyros and Glaucus torrents) have a small contribution in total water transfer into the Gulf of Patras, but may have an important contribution in solid waste transfer especially during heavy rain events.

Limassol Gulf in Cyprus is open to the Levantine Sea (Eastern Mediterranean Sea). The greater Limassol area counts approx. 192,000 inhabitants (2011, [www.mof.gov.cy/cystat](http://www.mof.gov.cy/cystat)). Marine navigation is considered intense (Limassol port 3430 vessel arrivals in 2013; [www.cpa.gov.cy](http://www.cpa.gov.cy)) while professional fisheries are moderate. Limassol Gulf receives the effects of various kinds of pressures (urbanization, tourism, commercial, industry, crafts, warehouses and aquaculture).

Constanta Bay is located in the NW Black Sea. It receives the impact of Constanta city, the second biggest city in Romania (425,916 inhabitants – INSSE 2011) and also the most important urban center for the whole Romanian coastal zone. The large protected deltaic zone, including the Danube Delta is located in the northern Romanian coastal zone. Marine navigation is considered extreme due to the presence of the port of Constanta (14,066 arrivals in 2013; [www.portofconstantza.com](http://www.portofconstantza.com)), the biggest port in the Black Sea. Professional and recreational fisheries are considered as intense.

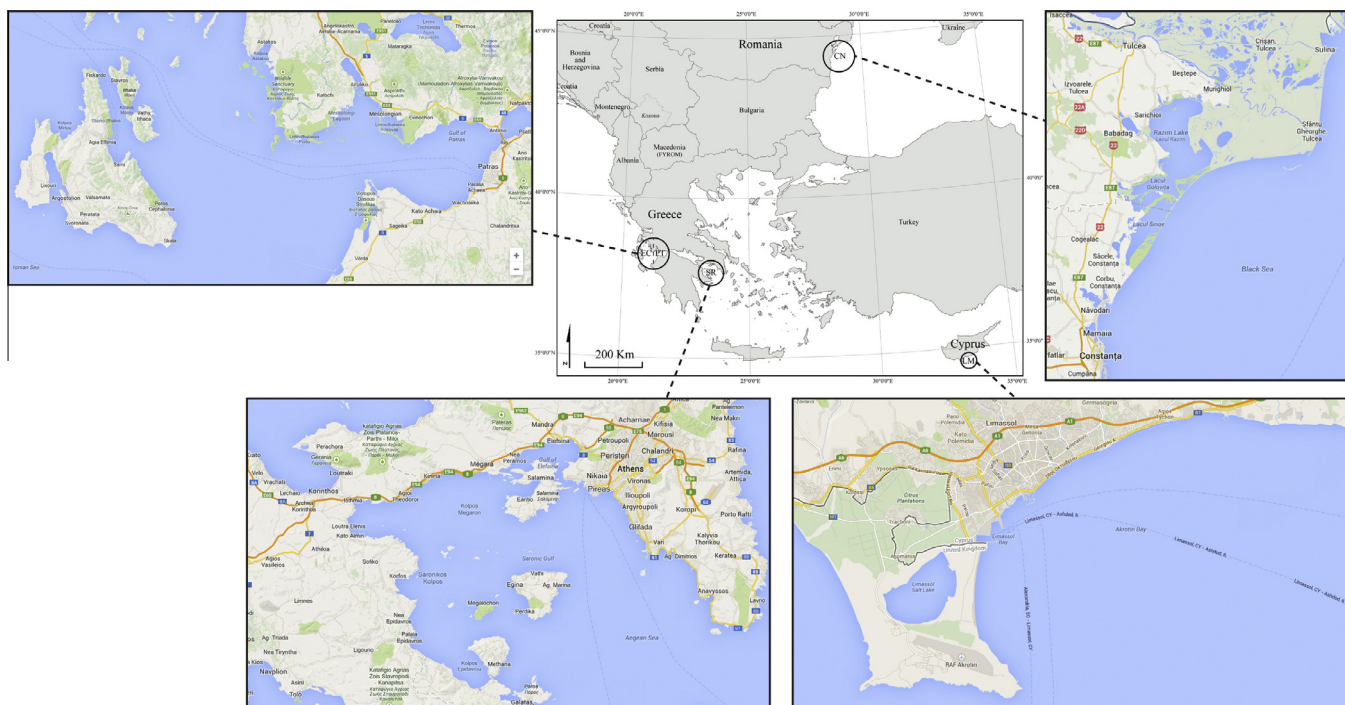
In the following Table 1, are summarized the pressures related to marine litter generating activities for the different study areas.

## 3. Methodology

### 3.1. Survey design

Benthic marine litter items were collected in the selected areas (SR, PT, EC, LM, CN) during the period January to March 2013. In the Mediterranean Sea (SR, PT, EC, LM) marine litter items were collected by professional trawlers. For each study-area, marine litter was collected by the same trawler in order to ensure data homogeneity and quality control. In the Black Sea (CN) marine litter items were collected by the R/V “Sea Star”. In Table 2, the characteristics of the vessels used in the surveys are summarized.

In total 94 trawlings were conducted in the studied areas (SR: 41; PT: 18; EC: 10; LM: 9; CN: 16) corresponding to a total of 713.6 km of surveyed area (SR: 207.1 km; PT: 192.6 km; EC: 125.7 km; LM: 112.3 km; CN: 75.9 km). All five areas are commonly used fishing grounds. Bottom trawlers with otter operate in Saronikos, Echinades and Limassol Gulfs 8 months per year, and in the Gulf of Patras 3 months per year. It should be mentioned here that sampling with bottom trawlers can only be applied in sandy or muddy substrates and not in rocky substrates.



**Fig. 1.** The five study areas: Saronikos Gulf (SR), Gulf of Patras (PT), Echinades Gulf (EC), Limassol Gulf (LM) in the Eastern Mediterranean Sea; Constanta Bay (CN) in the Black Sea.

### 3.2. Data collection

Marine litter items collected from each trawl were separately stored in bags and all the corresponding trawling characteristics were recorded (GPS positioning data, cruise speed, sac opening, date, place). Each bag was weighted and then marine litter items were sorted and classified according to the Master List of categories of the guidance document (TGML/JRC) (Galvani et al., 2013b). The basic material types are as follows: (i) plastic, (ii) metal, (iii) rubber, (iv) glass/ceramics, (v) natural products and (vi) miscellaneous with a total of thirty-nine (39) subcategories. Classification was also conducted according to six successive size categories: (i)  $<5 \times 5$  cm, (ii)  $<10 \times 10$  cm, (iii)  $<20 \times 20$  cm, (iv)  $<50 \times 50$  cm, (v)  $<100 \times 100$  cm, (vi)  $>100 \times 100$  cm. Each category refers to the surface area of each item, estimated visually. The small items ( $<5 \times 5$  cm), whose size was equal to the mesh size, were found entangled with other larger items. These small sized items were counted as part of the sample but it is very likely that are underestimated as more of them might be present on the sea floor and pass through the net. Naturally occurring items such as seaweed and driftwood particles were excluded from the recordings and differentiated from the *natural products* category, which actually refers to processed materials.

Density values (items/km<sup>2</sup>) were calculated using the total distance covered during the trawling (detailed GPS spotting) multiplied by the opening of the mouth of the trawling sac.

## 4. Results

### 4.1. Marine litter abundance and spatial distribution

In total 5398 marine litter items (Table 3) were collected from all the study areas (SR: 3269; PT: 1293; EC: 580; LM: 46; CN: 210). In the Saronikos Gulf were recorded on average 80 items/trawl (22–176 items/trawl), in the Gulf of Patras 72 items/trawl (6–202 items/trawl) and in the Echinades Gulf 58 items/trawl (6–101 items/trawl). On average, 5 items/trawl were recorded in

Limassol Gulf, two trawls were litter free and the rest of the recordings ranged from 1 to 18 items/trawl. Lastly, in Constanta Bay 13 items/trawl (1–48 items/trawl) were recorded. On average, 352 kg of marine litter were collected in Saronikos Gulf; 369 kg in the Gulf of Patras; 75 kg in the Echinades Gulf; 15 kg in the Limassol Gulf and 87 kg in Constanta Bay. These amounts correspond to: 2.5–26.8 kg/trawl in Saronikos Gulf, 2.1–56.5 kg/trawl in the Gulf of Patras, 1.6–24.5 kg/trawl in Echinades Gulf, 0.2–8.6 kg/trawl in Limassol Gulf with two trawls being litter-free and 0.8–23.5 kg/trawl in Constanta Bay.

Mean marine litter density was calculated for each study area. The highest density was found in the Saronikos Gulf ( $1211 \pm 594$  items/km<sup>2</sup>), almost half the density was found in the Gulf of Patras ( $641 \pm 579$  items/km<sup>2</sup>), followed by the Echinades Gulf ( $416 \pm 379$  items/km<sup>2</sup>), the Constanta Bay ( $291 \pm 237$  items/km<sup>2</sup>) and the Limassol Gulf ( $24 \pm 28$  items/km<sup>2</sup>). In Fig. 2, trawling transects are plotted according to marine litter density (items/km<sup>2</sup>); so that the thickness of the transect lines is proportional to marine litter density (items/km<sup>2</sup>). In this way marine litter accumulation can be visualized with detail relatively to its spatial distribution and the corresponding trawling length. Although the trawling samplings were concentrated on the fishing grounds on the gulfs and thus not distributed randomly, some general spatial trends can be inferred. In the Saronikos Gulf high marine litter average densities were equally distributed at all sampled isobaths (50 m: 1145 items/km<sup>2</sup>; 200 m: 1096 items/km<sup>2</sup>; 250 m: 1423 items/km<sup>2</sup>; 300 m: 979 items/km<sup>2</sup>; 350 m: 1182 items/km<sup>2</sup>) without any particular distribution pattern. The variation observed in the thickness of the lines (Fig. 2a) in the Saronikos Gulf is related to the fact that in many cases trawlings were conducted at the same area within one-day time lag. The majority of litter was collected during the first samplings while the short time lag did not permit the accumulation of 'new' litter.

In the Echinades Gulf the highest densities (880 items/km<sup>2</sup> on average) were recorded close to the coast and the Achelous River mouth at depths of about 50 m, falling to approximately 300 items/km<sup>2</sup> at the 100 m, 150 m and 200 m isobaths. In the Gulf

**Table 1**  
Summary of the various marine litter generation activities for the five study areas (SR, PT, EC, LM, CN).

Study area	Population (inhabitants)	Ports/harbors	Marine traffic (arrivals/yr)	Marinas, small ports, fishing shelters	Number of professional fishing vessels	River runoff (m <sup>3</sup> /yr)	Gulf geomorphology
SR	4,000,000	1	17,525	39	1111	N/A*	Semi-enclosed
PT	250,000	1	1923	7	400	1000 × 10 <sup>6</sup>	Semi-enclosed
EC	50,000	0	1,923	21	400	7800 × 10 <sup>6</sup>	Open
LM	101,000	1	3430	4	365	11 × 10 <sup>6</sup>	Open
CN	425,916	1	14,066	3	439	199,000 × 10 <sup>6</sup>	Open

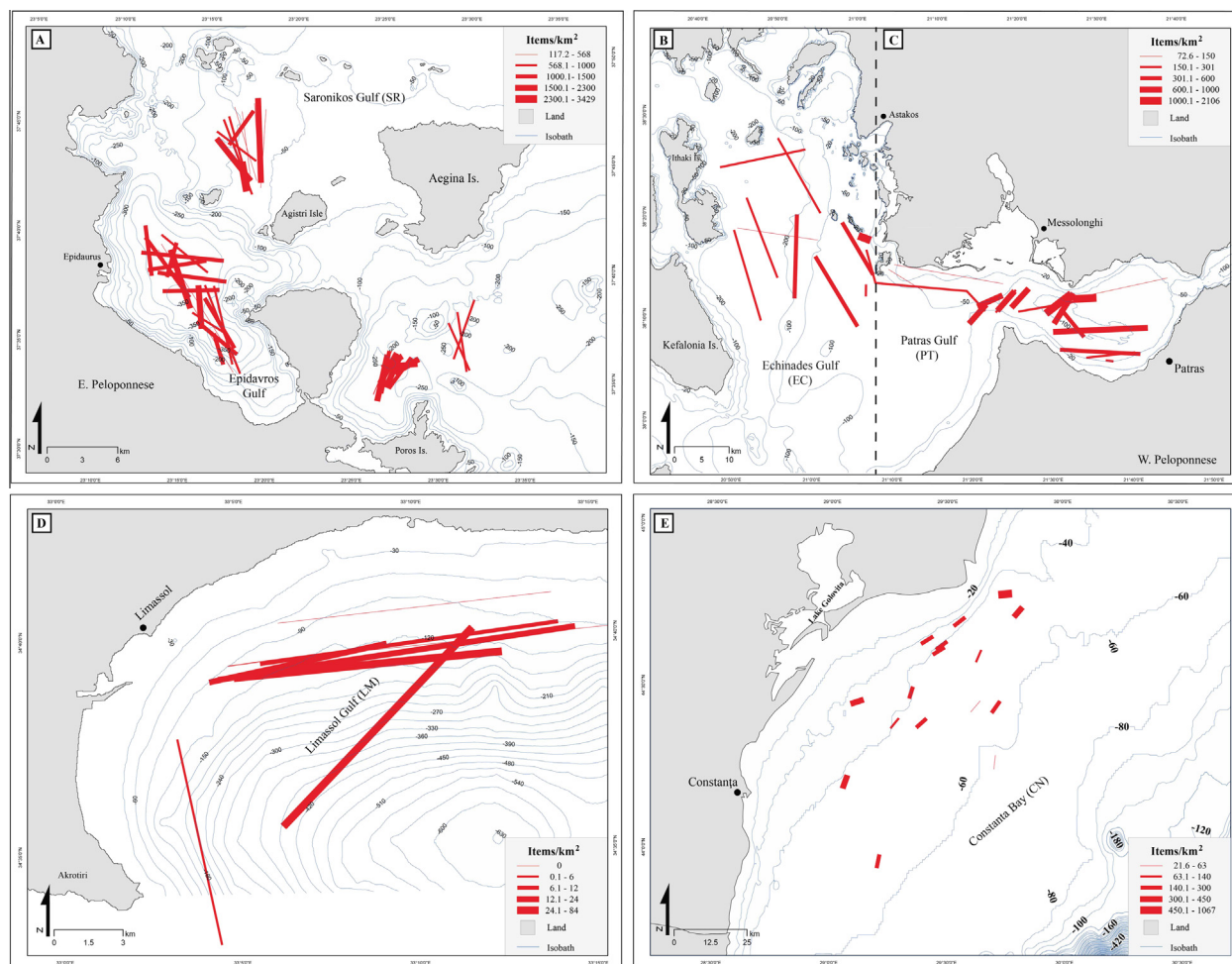
\* Not available published data.

**Table 2**  
Vessel characteristics used for benthic litter surveys.

	Greece Saronikos Gulf	Greece Gulf of Patras	Greece Echinades Gulf	Cyprus Limassol Gulf	Romania Constanta Bay
Vessel type	Prof. Trawler	Prof. Trawler	Prof. Trawler	Prof. Trawler	R/V Vessel
Vessel length	16.5 m	20 m	20 m	22 m	25.8 m
Trawl speed	2.2–2.3 knots	2.5–3.5 knots	2.5–3.5 knots	2.6–3.0 knots	2.5 knots
Sac Opening	15 m	15 m	15 m	15 m	9.5 m
Mesh Opening	50 mm	50 mm	50 mm	50 mm	20 mm

**Table 3**  
Composition (%) of marine litter in the five study areas.

%	Saronikos Gulf (SR)	Gulf of Patras (PT)	Echinades Gulf (EC)	Limassol Gulf (LM)	Constanta Bay (CN)
A. Plastics	95.0 ± 11.9	59.9 ± 9.2	67.4 ± 7.7	67.4 ± 14.4	45.2 ± 4.8
A1. Bottles	10.1	8.9	16.5	6.5	14.3
A2. Sheets	31.1	5.3	3.3	2.2	5.2
A3. Bags	36.5	34.0	27.1	54.3	12.9
A4. Caps/lids	0.2	0.0	0.9	0.0	2.4
A5. Fishing line (monofilament)	0.3	0.9	0.7	0.0	5.2
A6. Fishing line (entangled)	5.0	4.9	5.9	0.0	0.0
A7. Synthetic rope	5.4	0.5	0.2	0.0	1.9
A8. Fishing net	3.9	1.2	0.5	0.0	2.4
A9. Cable ties	0.1	0.0	0.0	0.0	0.0
A10. Strapping bands	0.0	0.0	0.0	0.0	0.0
A11. Crates/containers	0.0	3.8	11.0	0.0	0.5
A12. Diapers	0.0	0.0	0.0	0.0	0.0
A13. Sanitary towel/tampon	0.0	0.0	0.0	0.0	0.5
A14. Other...	2.4	0.5	1.4	4.3	0.0
B. Metals	3.1 ± 1.0	12.5 ± 4.4	20.5 ± 7.2	8.7 ± 2.3	21.9 ± 2.3
B1. Cans (food)	0.0	0.1	0.0	2.2	6.7
B2. Cans (beverage)	2.8	12.4	20.3	6.5	3.3
B3. Fishing related	0.03	0.0	0.0	0.0	2.4
B4. Drums	0.03	0.0	0.2	0.0	0.0
B5. Appliances	0.0	0.0	0.0	0.0	3.3
B6. Car parts	0.0	0.0	0.0	0.0	1.4
B7. Cables	0.1	0.0	0.0	0.0	4.8
B8. Other...	0.2	0.1	0.0	0.0	0.0
C. Rubber	0.1 ± 0.02	4.6 ± 1.1	2.6 ± 0.9	2.2 ± 1.0	1.4 ± 0.4
C1. Boots	0.03	0.3	0.2	0.0	0.5
C2. Balloons (boat)	0.0	0.5	0.0	0.0	0.0
C3. Bobbins (fishing)	0.03	0.2	0.2	0.0	1.0
C4. Tyre	0.0	0.7	0.2	2.2	0.0
C5. Other...	0.0	2.8	2.1	0.0	0.0
D. Glass/ceramics	0.1 ± 0.02	15.8 ± 6.0	3.8 ± 1.7	6.5 ± 3.3	22.4 ± 4.2
D1. Jar	0.0	0.9	0.3	0.0	4.8
D2. Bottle	0.03	12.9	3.4	6.5	8.6
D3. Pieces	0.0	2.0	0.0	0.0	9.0
D4. Other...	0.03	0.0	0.0	0.0	0.0
E. Natural products	0.1 ± 0.04	0.2 ± 0.04	0.5 ± 0.2	4.3 ± 1.2	7.1 ± 3.2
E1. Wood (processed)	0.1	0.1	0.0	2.2	7.1
E2. Rope	0.0	0.1	0.2	0.0	0.0
E3. Paper/cardboard	0.03	0.0	0.0	0.0	0.0
E4. Pallets (wooden)	0.0	0.0	0.0	0.0	0.0
E5. Other...	0.0	0.0	0.3	2.2	0.0
F. Miscellaneous	1.6 ± 0.9	7.04 ± 1.93	5.2 ± 1.6	10.9 ± 2.5	1.9 ± 1.1
F1. Clothing/rags	1.5	4.3	3.3	2.2	0.0
F2. Shoes	0.1	0.4	0.2	2.2	1.9
F3. Other...	0.03	2.4	1.7	6.5	0.0
Number of Trawls (n)	41	18	10	9	16



**Fig. 2.** Marine Litter density (items/km<sup>2</sup>) in Saronikos Gulf (A), Echinades Gulf (B), Gulf of Patras (C), Limassol Gulf (D), Constanta Bay (E). Line positioning correspond to the trawling transect; line thickness to marine litter density.

of Patras a reverse pattern can be observed with relatively lower densities (average 440 items/km<sup>2</sup>) close to the city of Patras at the 50 m isobath and almost twice higher (average 893 items/km<sup>2</sup>) in the deeper parts of the gulf (100 m depth). Also in the Limassol Gulf, the highest litter densities, reaching 84 items/km<sup>2</sup>, were found in the deeper parts. Finally, in Constanta Bay we recorded the highest density (1068 items/km<sup>2</sup>) in front of the Danube mouth, whereas no consistent distribution pattern in relation to depth was observed since all the samplings were conducted on the shelf (at depths <60 m).

In the same rational, the densities (items/km<sup>2</sup>) of plastics (Fig. 3) and metals (Fig. 4) were plotted for the study areas. Plastics were predominant in all study areas; the highest densities were observed in Saronikos Gulf, where high densities were found in all transects. In the Gulf of Patras the highest densities of plastics were observed in transects that were adjacent to the city of Patras. In Limassol Gulf the high plastic densities were observed in the deepest transects while in Constanta Bay the highest densities were observed at two transects closer to the Danube Delta.

Metals in Saronikos Gulf and the Constanta Bay, seem to accumulate in the deeper parts of the surveyed areas while in the Gulf of Patras the highest accumulation of metals was observed in the northern part and in the Echinades Gulf, close to the Acheloos and Evinos rivers estuaries. In the Limassol Gulf the metals, like the plastics, accumulated in the deeper part of the gulf.

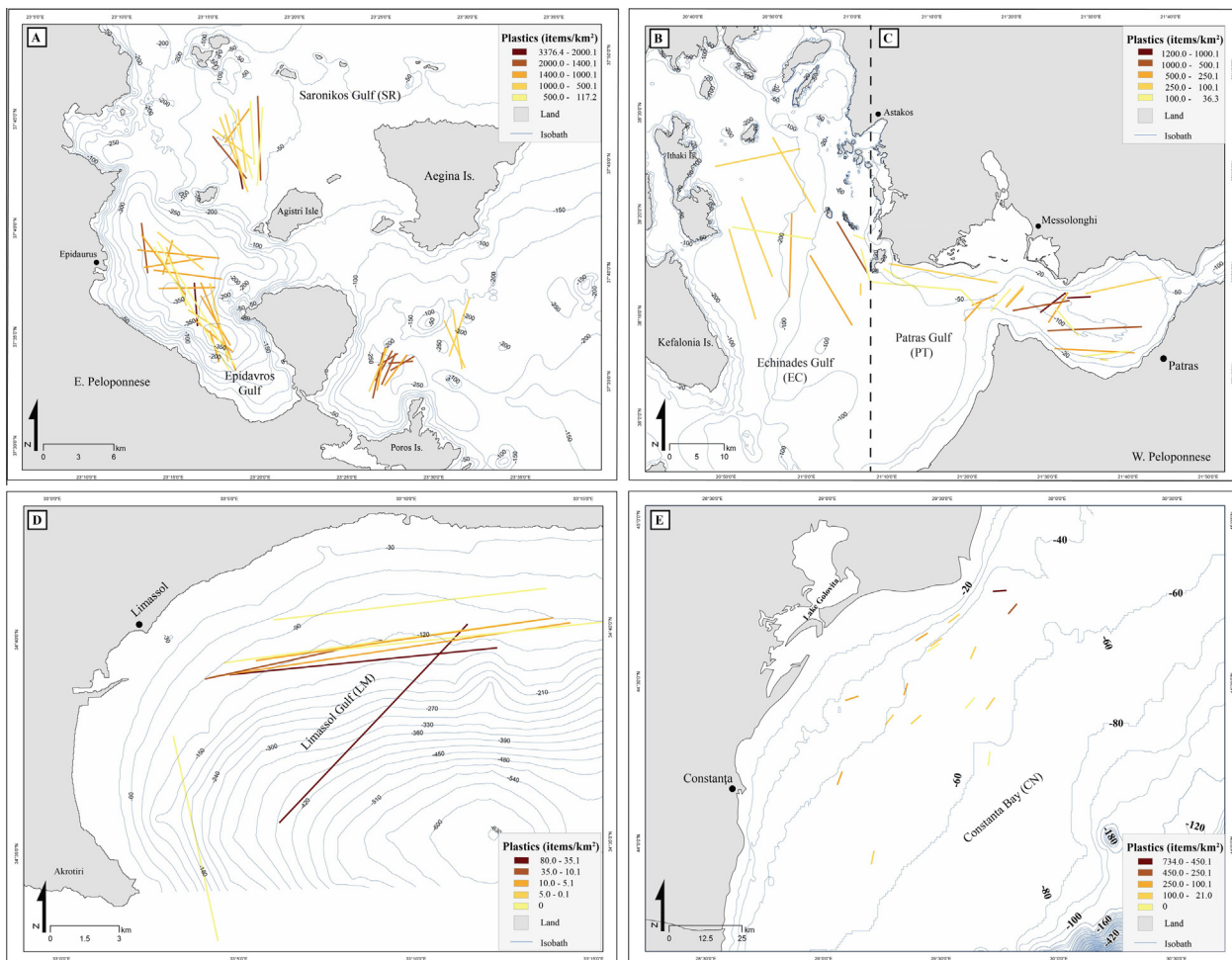
#### 4.2. Marine litter composition

In Table 3 the percentage contribution of each marine litter category of the Master List (TGML/JRC) is presented for the five studied areas.

Plastics were the most abundant marine litter type in all study areas (SR: 95%, EC: 67%, PT: 60%, LM: 67%, CN: 45%). Especially in the Saronikos Gulf *plastics* contribution was dominant (95%), shrinking the percentages of all the remaining litter categories to ≤3%. The following two most abundant marine litter types were *Metals* (EC: 20%, PT: 12%, LM: 8.7%, CN: 22%) and *Glass/Ceramics* (EC: 6%, PT: 16%, LM: 7%, CN: 22%). *Rubbers, natural products* and *miscellaneous* categories contributed with low percentages (<10%).

Within *plastics*; *bags, bottles* and *sheets* had the highest contribution for all the study areas with mean abundance of 49.6%, 17.5% and 13.5% respectively. *Fishing lines* had an overall contribution of 6.7% with both monofilament (3.3%) and entangled (3.4%) having same share, whereas *fishing net* (2.8%) and *synthetic rope* (2.6%) were the less abundant items.

Within *metals*; *beverage cans* (69.8%) and *food cans* (13.9%) had the highest share. Especially for the Saronikos Gulf, Echinades Gulf and the Gulf of Patras they exceeded 90% of total metals. In Constanta Bay, *cables* exceeded 21% while the fishing related metal items had a significant contribution (10.9%).



**Fig. 3.** Densities (items/km<sup>2</sup>) of plastic marine litter in Saronikos Gulf (A), Echinades Gulf (B), Gulf of Patras (C), Limassol Gulf (D) and Constanta Bay (E). Line positioning corresponds to the trawling transect and the color scale to the density range. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

#### 4.3. Marine litter sizing

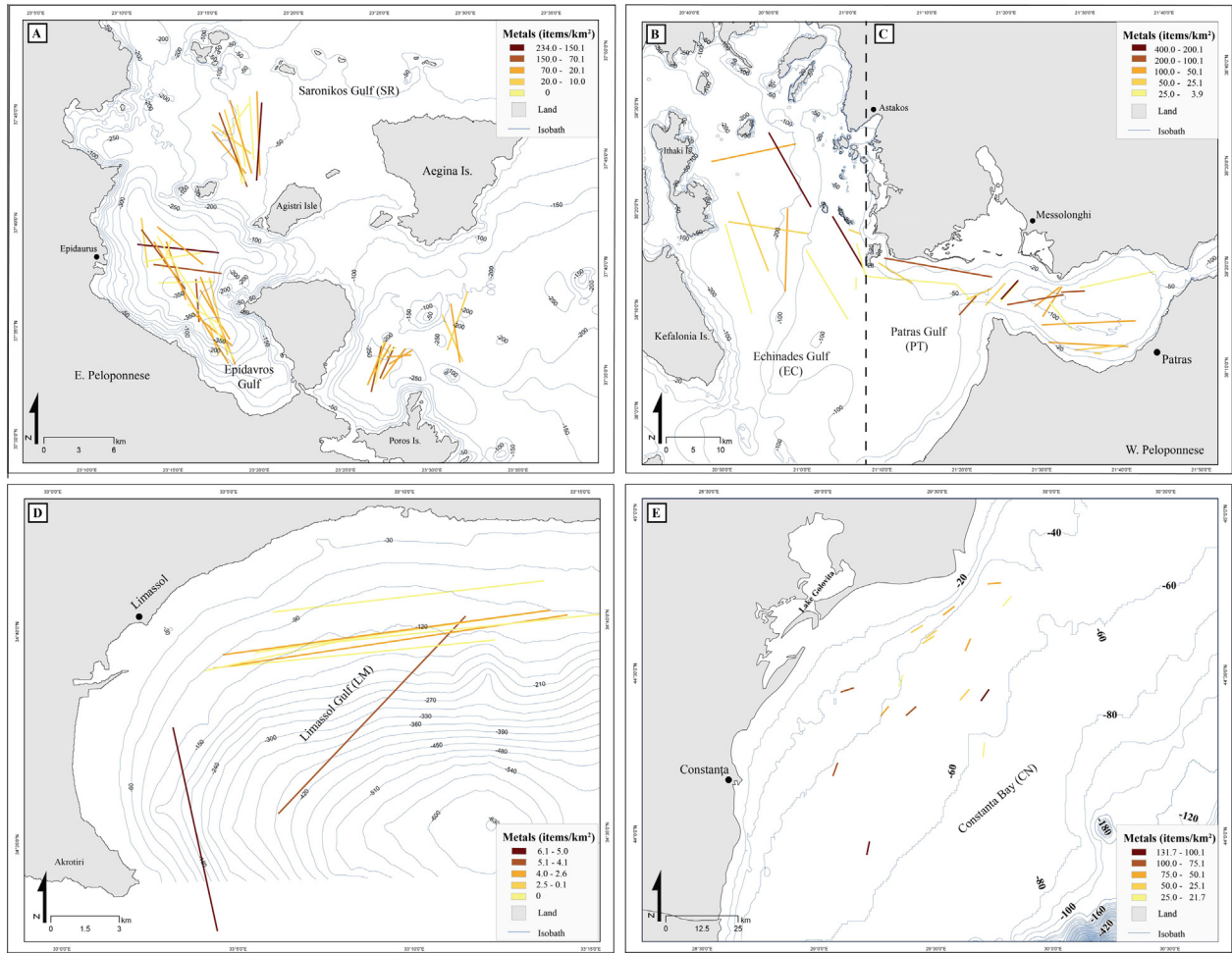
Another parameter that was investigated was the variation of size distribution among the different marine litter items. In many cases, the comparison of density values (items/km<sup>2</sup>) with litter weight (kg) cannot be done with success, as this can be highly dependent on marine litter size variation. Fig. 5 shows the size distribution of the benthic litter items in the study areas. In all the areas more than half of the recorded marine litter items were classified into medium size categories i.e. <10 × 10 cm and <20 × 20 cm (SR: 67.8%, PT: 50.8%, LM: 71.8%, CN: 66.2%) while big items (<100 × 100 cm, >100 × 100 cm) had a relatively small share (SR: 0.9%, PT: 5.3%, LM: 0.0%, CN: 3.3%). The most striking finding was that small sized (<5 × 5 cm) items had a significant contribution in all the study areas (SR: 23.0%, PT: 6%, EC: 7% LM: 13%, CN: 22%). The presence of significant amount of small sized items shows the importance of marine litter fragmentation. Especially for plastics, there seems to be an important stock of plastic fragments on the sea bottom with yet undefined effect on marine environment and marine biota.

#### 5. Discussion

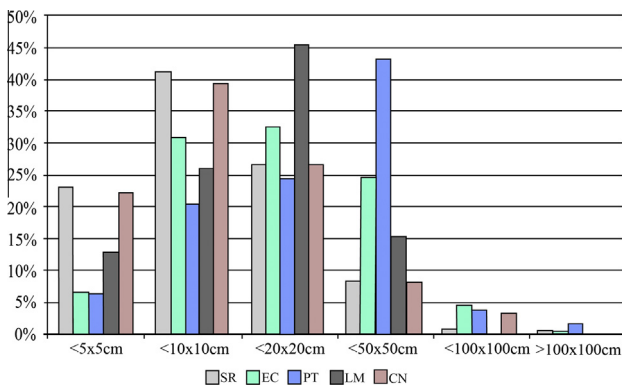
Marine Strategy Framework Directive (MSFD) includes marine litter as one of the eleven qualitative descriptors that have to be considered towards achieving GES. During the last decade we

might have compensated the “lack of knowledge” in regard to the problem of marine litter but we still have way to cover in order to fulfill the “lack of data”. The increased variability found in litter abundance and distribution within each study area as well as among all five areas suggests that several factors affect marine litter accumulation in the benthic environments. Furthermore this observation underlines the importance of systematic monitoring programs in order to acquire ‘sound’ information for setting the targets to GES.

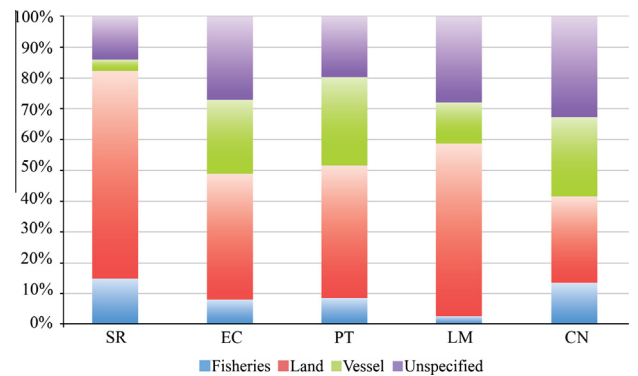
The marine litter load for each survey area is expected to be affected by a series of factors including; (i) the proximity of big cities to our study areas, (ii) the intensity of marine traffic, (iii) the riverine inputs into the sea and (iv) the geomorphology of the area (semi-enclosed vs open – shelf environment). Surprisingly, we did not find a direct relationship of marine litter density with population density (Leite et al., 2014). Although these two parameters showed the highest values in the Saronikos Gulf, in Constanta Bay which is the second most populated area, marine litter density was found lower than in the Gulf of Patras or the Echinades Gulf. Similarly, the increased marine traffic reported for the Saronikos Gulf, Constanta Bay and Limassol Gulf is not linearly reflected in the corresponding marine litter densities. Again the Gulf of Patras and Echinades Gulf where the lowest ship arrivals per year are recorded showed marine litter densities higher than those found in Constanta Bay or Limassol Gulf. Concerning the importance of riverine outflows on marine litter input, it is understood that



**Fig. 4.** Densities (items/km<sup>2</sup>) of metal marine litter items in Saronikos Gulf (A), Echinades Gulf (B), Gulf of Patras (C), Limassol Gulf (D) and Constanta Bay (E). Line positioning correspond to the trawling transect; and the color scale to the density range. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)



**Fig. 5.** Size distribution of marine litter items in the five study areas: Saronikos Gulf (SR), Gulf of Patras (PT), Echinades Gulf (EC), Limassol Gulf (LM) and Constanta Bay (CN).



**Fig. 6.** The different sources (fisheries, land, vessel, unspecified) of marine litter items in the study areas.

Constanta Bay has by far the greatest influence as it is situated in the vicinity of the Danube River delta. Nevertheless, our data from the five study areas did not reflect the importance of Danube River as source of marine litter; in particular the litter density close to the Danube river estuary was found comparable to those in Echinades Gulf which is influenced by the Acheloos river outflow or Saronikos Gulf and the Gulf of Patras. The fact that we recorded

the maximum litter density, for this area, in front of the Danube river mouth as well as some large metallic items such as cables, car parts and appliances give us a hint that substantial amounts of marine litter from land-based sources, transferred through the Danube River are not accumulated on the shelf area. From the above it becomes clear that the external sources of litter are not directly reflected on the benthic litter densities found in the five studied areas, and transport processes taking place within the

marine environment seem to be equally important for marine litter redistribution on the sea bottom. For places such as the Gulf of Lions, the Bay of Biscay and Monterey Bay (Galgani et al., 2000; Watters et al., 2010; Schlining et al., 2013) it has been found that marine litter is transported farthest from the shelf areas and accumulated in the deep canyons. It is possible that a similar situation occurs in the Limassol Gulf where the deeper parts show somewhat higher litter densities. Unfortunately, in Constanta Bay the maximum sampling depths did not exceed 60 m so we cannot test this hypothesis. As already mentioned the Gulf of Patras and Echinades Gulf show marine litter densities higher than expected based on the population density, marine traffic and riverine outflows. We believe that in this case the enclosed character of the gulfs and the relatively steep bottom topography favor the accumulation of litter there. This could be also the case for the Saronikos Gulf, where only the deepest western basin of the gulf was sampled. According to Kontoyiannis (2010) an anti-cyclonic surface water circulation is dominant most of the year, so it is possible that surface litter is transferred towards the western basin where it is trapped and eventually loses its buoyancy through entanglement and fouling and settles. So, the maximum litter densities recorded in the western basin of the Saronikos Gulf could be due to the combined effect of important sources (such as population, tourism and marine traffic) and transport processes.

A way to discern the importance of direct litter sources vs transport mechanisms is to look for the specific types of litter settled in a certain location. Clear linking is not always possible and in many cases it is very subjective. Nevertheless, there are certain marine litter types that linking to sources can be objective, i.e. fisheries related items (monofilament and entangled fishing lines, synthetic ropes, fishing nets, fishing gear, boots, bobbins and tyres re-used as fishing boat fenders). Among our study areas, the Saronikos Gulf had the highest link to fisheries marine litter sources (14.8%) followed by Constanta Bay (13.3%), the Gulf of Patras (8.7%), Echinades Gulf (7.8%) and Limassol Gulf (2.2%) (Fig. 6). These findings are in accordance with the number of professional fishing boats recorded for the studied areas (Table 1). Fisheries related items were relatively low in the present study compared to the high average of 40% or the extreme of 65% in the Celtic Shelf (Galgani et al., 2000). Some researchers (Horsman, 1982; Moore and Allen, 2000; Stafatos et al., 1999; Koutsodendris et al., 2008) have suggested that metal beverage cans and glass bottles indicate vessel-based sources because these types of litter are not expected to travel long distances so they are probably deposited close to the area of sampling. Derraik (2002) and Koutsodendris et al. (2008) have suggested that garbage bags, shopping bags, plastic sheets and crates/containers originate mainly from land-based sources including domestic, agricultural and industrial activities. During our surveys, metallic litter corresponded mainly to aluminum beverage cans and glass litter to bottles and jars while plastic litter mainly consisted of bags, sheets and crates/containers. Plastic bottles, clothing items, boots and the subcategory “others” are not indicator items since they could have originated both from land- and marine-based sources. Based on the above mentioned marine litter categories which are only indicative to certain litter sources, land-based sources (domestic, agricultural and industrial activities) contributed to 67.6%, 43.1%, 41.1%, 56.5% and 28.1% of the total litter collected in the Saronikos Gulf, the Gulf of Patras, the Echinades Gulf, the Limassol Gulf and the Constanta Bay, respectively (Table 3, Fig. 6). Lower percentages of litter (SR: 3.1%, PT: 28.2%, EC: 24%, LM: 13%, CN: 25.7%) were attributed to vessel-based sources with the exception of Constanta Bay where the percentage of the land originated and the vessel originated litter was almost the same (Table 3, Fig. 6). A significant percentage of marine litter items could not be directly linked to a particular source and hence it could be attributed to both sources (land- and vessel based) in

the survey areas (SR: 14.5%, PT: 20.0%, EC: 27.1%, LM: 28.3%, CN: 32.9%) (Table 3, Fig. 6).

The very low contribution of items associated with the marine sources (metallic beverage cans and glass bottles) found in the Saronikos Gulf (3.1%) is due to the fact that our sampling area (western basin) does not coincide with the main navigation routes, so most litter found there was transported from land based sources (67%) as already explained previously. A more complete picture of litter abundance in the Saronikos Gulf will be obtained during future surveys covering more locations. For all the other areas the types of litter items found reflect substantial contribution from both land and marine-based human activities.

Plastics were, by far, the most abundant marine litter type in all studied sites while in Saronikos Gulf (95%) they exceeded the global average of 75% (Galgani et al., 2013a). Four litter items (plastic bags, plastic bottles, glass bottles and metal beverage cans) accounted for over 40% of total items for the survey areas. This is in accordance with the findings of an EU-wide survey in which plastic bags and bottles accounted up to 70% of total debris in the Baltic Sea, the North Sea, the Celtic Sea, the Bay of Biscay and different areas in the north-western basin of the Mediterranean Sea and the Adriatic Sea (Galgani et al., 2000). Over 8 billion plastic carrier bags were littered in the EU in 2010, representing one of the most important threats for the marine environment. The findings of the present survey further support the EU decision for discouraging the use of plastic carrier bags. Furthermore legislation to reduce the use of plastic carrier bags in countries of the Eastern Mediterranean and the Black Sea is strongly suggested.

Size variation for marine litter items was also investigated in order to get some information on the extent of litter fragmentation. Small sized items (<5 × 5 cm), which were entangled into bigger marine litter items, had significant proportion in all survey areas (SR: 23%, PT: 6%, EC: 7%, LM: 13%, CN: 22%). Visual inspection of small-sized litter showed that the majority of them are plastic fragments that have been formed by the physical, chemical and biological fragmentation of larger items and not from the direct release of small sized plastic items. The significant amount of small-sized plastic litter in the five study areas coincides well with the increase of meso- and microplastic litter in global ocean. The UK Marine Conservation Society (MCS, 2007) reports a 30% increase in the abundance of larger fragments (1–50 cm in size) followed by a 20% increase in the abundance of smaller fragments (<1 cm) between 1998 and 2006. Thompson et al. (2004) found that there is a significant increase of microplastic abundance over the last 40 years. The elevated percentage of small plastic fragments in the Eastern Mediterranean Sea and the Black Sea is of particular concern because they are difficult to be removed from the marine environment, they have the potential to be easily ingested by organisms and they transfer harmful chemicals to living organisms (Barnes et al., 2009).

Benthic marine litter surveys concerning the trends in litter accumulation and the temporal variation of litter abundance are very rare and the available data have shown high variability (Barnes et al., 2009). Two of the few areas worldwide where temporal data is available on benthic marine litter over a period of 15 years are Patras and Echinades Gulfs (Stafatos et al., 1999; Koutsodendris et al., 2008). Back in 1997–1998, marine litter density in Patras Gulf was found 240 items/km<sup>2</sup> based only on two trawlings and in Echinades Gulf was 89 items/km<sup>2</sup> (Stafatos et al., 1999), while for the period 2000–2003 30 trawlings gave an average marine litter density of 313 items/km<sup>2</sup> (Koutsodendris et al., 2008). The present study, clearly demonstrates a significant increase in benthic marine litter abundance in Patras Gulf (641 items/km<sup>2</sup>). Albeit the increase of marine litter abundance, typology seems to follow more or less similar patterns especially from 2000 onwards. This temporal increase of litter abundance in



the two Gulfs is in contradiction to temporal trends that have been observed in other areas. On the sea-bed of Tokyo Bay, litter densities decreased from 1996 to 2003 (Kanehiro et al., 1995; Kuriyama et al., 2003) while in the Gulf of Lions litter abundance slightly decreased during 1994–2009 (Galgani et al., 2013a).

The interpretation of the spatial and temporal variations of marine litter is a very challenging issue since the litter distribution is influenced by a variety of factors i.e. human activities, seasonal changes in rivers flow rate, hydrodynamic conditions and seafloor morphology as well as socioeconomic measures implemented in the different regions. Our results demonstrate the complexity of the problem and highlight the need for concentrated efforts in order to collect sufficient data to fill in the existing gaps in knowledge. In that aspect, the present work provides support to decision makers for the implementation of the MSFD, which will lead to the reduction of marine litter accumulation in the susceptible benthic marine environment.

### Acknowledgements

The authors would like to thank: (i) all participants in cruises for their assistance in data collection, (ii) Dr. Stavros-Nikiforos Spyrellis (UMR Géographie-cités, Equipe P.A.R.I.S), Mrs Paraskevi Drakopoulou (HCMR) and Mrs Hara Kyriakidou (HCMR) for their valuable contribution in mapping, (iii) Mr. Stavros Stylianou of the Cyprus Oceanography Center for providing bathymetry data.

The present study was conducted within the framework of the European IP Project FP7 – PERSEUS EC Grant Agreement No 287,600. C.I. has received a Ph.D. grant.

### References

- Adimey, N.M., Hudak, C.A., Powell, J.R., Bassos-Hull, K., Foley, A., Farmer, N.A., White, L., Minch, K., 2014. Fishery gear interactions from stranded bottlenose dolphins, Florida manatees and sea turtles in Florida, U.S.A. *Mar. Pollut. Bull.* 81 (1), 103–115.
- Anastasopoulou, A., Mytilineou, C., Smith, C.J., Papadopoulou, K.N., 2013. Plastic debris ingested by deep-water fish of the Ionian Sea (Eastern Mediterranean). *Deep-Sea Res.* 174, 11–13.
- Anderson, J.A., Alford, A.B., 2014. Ghost fishing activity in derelict blue crab traps in Louisiana. *Mar. Pollut. Bull.* 79 (1–2), 261–267.
- Bakir, A., Rowland, S.J., Thompson, R.C., 2014. Transport of persistent organic pollutants by microplastics in estuarine conditions. *Estuarine Coast. Shelf Sci.* 140, 14–21.
- Barnes, D.K.A., Galgani, F., Thompson, R.C., Barlaz, M., 2009. Accumulation and fragmentation of plastic debris in global environments. *Philos. Trans. R. Soc. B Biol. Sci.* 364 (1526), 1985–1998.
- Bond, A.L., Provencher, J.F., Elliot, R.D., Ryan, P.C., Rowe, S., Jones, I.L., Robertson, G.J., Wilhelm, S.L., 2013. Ingestion of plastic marine debris by Common and Thick-billed Murres in the northwestern Atlantic from 1985 to 2012. *Mar. Pollut. Bull.* 77, 192–195.
- Derraik, J.C.B., 2002. The pollution of the marine environment by plastic debris: a review. *Mar. Pollut. Bull.* 44, 842–852.
- Galgani, F., Jaunet, S., Campillo, A., Guenegen, X., His, E., 1995. Distribution and abundance of debris on the continental shelf of the North-Western Mediterranean Sea. *Mar. Pollut. Bull.* 30, 713–717.
- Galgani, F., Souplet, A., Cadiou, Y., 1996. Accumulation of debris on the deep sea floor off the French Mediterranean coast. *Mar. Ecol. Prog. Series* 142, 225–234.
- Galgani, F., Leaute, J.P., Moguelet, P., Souplet, A., Verin, Y., Carpentier, A., Goraguer, H., Latrouite, D., Andral, B., Cadiou, Y., Mahe, C., Poulard, J.C., Nerisson, P., 2000. Litter on the sea floor along European Coasts. *Mar. Pollut. Bull.* 40 (6), 516–527.
- Galgani, F., Hanke, G., Werner, S., De Vrees, L., 2013a. Marine litter within the European marine strategy framework directive. *ICES J. Mar. Sci.* 70, 1055–1064.
- Galgani, F., Hanke, G., Werner, S., Oosterbaan, L., Nilsson, P., Fleet, D., Kinsey, S., Thompson R.C., Van Franeker, J., Vlachogianni, T., Scoullou, M., Mira Veiga, J., Palatinus, A., Matiddi, M., Maes, T., Korpinen, S., Budziak, A., Leslie, H., Gago, J., Liebezeit, G., 2013b. Guidance on Monitoring of Marine Litter in European Seas. MSFD Technical Subgroup on Marine Litter (TSG-ML).
- Galiil, B.S., Golik, A., Turkay, M., 1995. Litter at the bottom of the sea: a sea bed survey in the Eastern Mediterranean. *Mar. Pollut. Bull.* 30 (1), 22–24.
- Goldstein, M.C., Goodwin, D.S., 2013. Gooseneck barnacles (*Lepas spp.*) ingest microplastic debris in the North Pacific Subtropical Gyre. *PeerJ* 1, 184. <http://dx.doi.org/10.7717/peerj.184>.
- Horsman, P.V., 1982. The amount of garbage pollution from merchant ships. *Mar. Pollut. Bull.* 13, 167–169.
- Ivar do Sul, J.A., Costa, M.F., 2014. The present and future of microplastic pollution in the marine environment. *Environ. Pollut.* 185, 352–364.
- Kanehiro, H., Tokai, T., Matuda, K., 1995. Marine litter composition and distribution on the sea-bed of Tokyo Bay. *Fish. Eng.* 31, 195–199.
- Karapanagioti, H.K., Klontza, I., 2007. Investigating the properties of plastic resin pellets found in the coastal areas of Lesbos island. *Global NEST J.* 9 (1), 71–76.
- Karapanagioti, H.K., Klontza, I., 2008. Testing phenanthrene distribution properties of virgin plastic pellets and plastic eroded pellets found on Lesbos island beaches (Greece). *Mar. Environ. Res.* 65, 283–290.
- Karapanagioti, H.K., Endo, S., Ogata, Y., Takada, H., 2011. Diffuse pollution by persistent organic pollutants as measured in plastic pellets sampled from various beaches in Greece. *Mar. Pollut. Bull.* 62, 312–317.
- Katsanevakis, S., Katsarou, A., 2004. Influences on the distribution of marine debris on the seafloor of shallow coastal areas in Greece (Eastern Mediterranean). *Water Air Soil Pollut.* 158, 325–337.
- Katsanevakis, S., Verriopoulos, G., Nicolaidou, A., Thessalou-Legaki, M., 2007. Effect of marine litter on the benthic megafauna of coastal soft bottoms: a manipulative field experiment. *Mar. Pollut. Bull.* 54, 771–778.
- Keller, A.A., Fruh, E.L., Johnson, M.M., Simon, V., McGourty, C., 2010. Distribution and abundance of anthropogenic marine debris along the shelf and slope of the US West Coast. *Mar. Pollut. Bull.* 60, 692–700.
- Kontoyiannis, H., 2010. Observations on the circulation of the Saronikos Gulf: a Mediterranean embayment sea border of Athens, Greece. *J. Geophys. Res.* 115, C06029. <http://dx.doi.org/10.1029/2008JC005026>.
- Koutsodendris, A., Papatheodorou, A., Kougiourouki, O., Georgiadis, M., 2008. Benthic marine litter in four Gulfs in Greece, Eastern Mediterranean: abundance, composition and source identification. *Estuarine Coast. Shelf Sci.* 77, 501–512.
- Kuriyama, Y., Tokai, T., Tabata, K., Kanehiro, H., 2003. Distribution and composition of litter on seabed of Tokyo Gulf and its age analysis. *Nippon Suisan Gakkaishi* 69, 770–781.
- Leite, A.S., Santos, L.L., Costa, Y., Hatje, V., 2014. Influence of proximity to an urban center in the pattern of contamination by marine debris. *Mar. Pollut. Bull.* 81 (1), 242–247.
- Lopez-Carillo, L., Hernandez-Ramirez, R.U., Calafat, A.M., Torres-Sanchez, L., Galvan-Portillo, M., Needham, L.L., Ruiz-Ramos, R., Cebrían, M.E., 2010. Exposure to phthalates and breast cancer risk in Northern Mexico. *Environ. Health Perspect.* 114 (4), 539–544.
- Lusher, A.L., McHugh, M., Thompson, R.C., 2013. Occurrence of microplastics in the gastrointestinal tract of pelagic and demersal fish from the English Channel. *Mar. Pollut. Bull.* 67, 94–99.
- Madeira Di Benedetto, A.P., Arruda Ramos, R.M., 2014. Marine debris ingestion by coastal dolphins: what drives differences between sympatric species? *Mar. Pollut. Bull.* 83, 298–301.
- Mathalon, A., Hill, P., 2014. Microplastic fibers in the intertidal ecosystem surrounding Halifax Harbor, Nova Scotia. *Mar. Pollut. Bull.* 81, 69–79.
- MCS, 2007. Beachwatch 2006—the 14th Annual Beach Litter Survey Report. Marine Conservation Society, Ross on Wye.
- Mizukawa, K., Takada, H., Ito, M., Geok, Y., Hosoda, J., Yamashita, R., Saha, M., Suzuki, S., Miguez, C., Frias, J., Cepeda Antunes, J., Sobral, P., Santos, I., Micaelo, C., Ferreira, A.M., 2013. Monitoring of a wide range of organic micropollutants on the Portuguese coast using plastic resin pellets. *Mar. Pollut. Bull.* 70, 296–302.
- Moore, S.L., Allen, M.J., 2000. Distribution of anthropogenic and natural debris on the Mainland Shelf of the Southern California Bight. *Mar. Pollut. Bull.* 40 (1), 83–88.
- Pham, C.K., Gomes-Pereira, J.N., Isidro, E.J., Santos, R.S., Morato, T., 2013. Abundance of litter on Condor seamount (Azores, Portugal, Northeast Atlantic). *Deep Sea Res. Part II: Trop. Stud. Oceanogr.*, 204–208.
- Ramirez-Llodra, E., De Mol, B., Company, J.B., Coll, M., Sardà, F., 2013. Effects of natural and anthropogenic processes in the distribution of marine litter in the deep Mediterranean Sea. *Prog. Oceanogr.* 118, 273–287.
- Sanchez, P., Mercedes, M., Saez, R., Juan, S., Muntadas, A., Demestre, M., 2013. Baseline study of the distribution of marine debris on soft-bottom habitats associated with trawling grounds in the northern Mediterranean. *Sci. Mar.* 77 (2), 247–255.
- Schlining, K., Von Thun, S., Kuhn, L., Schlining, B., Lundsten, L., Jacobsen Stout, N., Chaney, L., Connor, J., 2013. Debris in the deep: using a 22-year video annotation database to survey marine litter in Monterey Canyon, central California, USA. *Deep-Sea Res.* 179, 96–105.
- Stefatos, A., Charalampakis, M., Papatheodorou, G., Ferentinos, G., 1999. Marine debris on the seafloor of the Mediterranean Sea: examples from two enclosed gulfs in Western Greece. *Mar. Pollut. Bull.* 36, 389–393.
- Thompson, R.C., Olsen, Y., Mitchell, R.P., Davis, A., Rowland, S.J., John, A.W.G., McGonigle, D., Russell, A.E., 2004. Lost at sea: where is all the plastic? *Science* 304, 838.
- Topçu, E.N., Öztürk, B., 2010. Abundance and composition of solid waste materials on the western part of the Turkish Black Sea seabed. *Aquat. Ecosyst. Health Manage.* 13 (3), 301–306.
- UNEP, 2011. Assessment of the Status of Marine Litter, in the Mediterranean. United Nations Environmental Program, Athens.
- Waluda, C.M., Staniland, I.J., 2013. Entanglement of Antarctic fur seals at Bird Island, South Georgia. *Mar. Pollut. Bull.* 74, 244–252.
- Watters, D.L., Yoklavich, M.M., Love, M.S., Schroeder, D.M., 2010. Assessing marine debris in deep seafloor habitats off California. *Mar. Pollut. Bull.* 60, 131–138.
- World Bank, 2012. What a waste. A global review of solid waste management. Urban Developers Series Knowledge Paper (No 68135).